

## Ecological patterns of water bug (Hemiptera: Heteroptera) assemblages in karst springs: a case study from central Montenegro

by

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### Abstract

The composition of water bug communities from 32 springs located in the central part of Montenegro was investigated. Twenty five species were identified, including 13 reported as new to Montenegro. The most common species were *Hydrometra stagnorum* and *Velia* sp. (Gerromorpha). Our study in the central part of Montenegro revealed that environmental and faunistic classification of karstic springs based on water bug communities may not correspond with each other. According to environmental characteristics, springs were divided into three groups indicating anthropogenic impact on the spring habitats. Water bug communities divided springs into four groups. There are differences in species richness between these four types of water bug assemblages and among the studied spring types. Results of CCA analysis revealed spring size as the main driver of biotic diversity of aquatic bugs in springs. Our study showed that community groups of water bugs specified in the biotic classification of spring habitats are much better defined than assemblages distinguished in the environmental site classification.

**Key words:** water bugs, Montenegro, springs, crenobiology, diversity

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## Introduction

Water bugs are an important component of the aquatic fauna. They are widely distributed and inhabit aquatic and semi-aquatic habitats, including both lentic and lotic bodies of water (Souza et al. 2006). In general, water bugs have a high dispersion capacity (Savage 1994; Usseglio-Polatera et al. 2000; Wachmann et al. 2006). For this reason, it is difficult to define the types of water bug assemblages (Carbonell et al. 2011). Although found in all types of aquatic habitats, they do not prefer high water velocity and prevail in stagnant or slow-flowing waters (Karaouzas, Gritsalis 2006; Nosek et al. 2007; Skern et al. 2010). Generally, the dimensions of water bodies, land use, aquatic and riparian vegetation, and water chemistry were considered as the most important factors affecting the assemblages of water bugs (Hufnagel et al. 1999; Karaouzas, Gritsalis 2006). Little is known about the ecology of water bugs in springs and there are no research on ecological factors that determine species distribution in these habitats.

In general, the fauna of water bugs in Montenegro is insufficiently studied (Schumacher 1914; Horváth 1918; Novak, Wagner 1955; Filippi 1957; Grupče 1961; Wagner 1962; Štusák 1980; Protić et al. 1990; Protić 1998; Kment et al. 2005; Kovács et al. 2011; Aukema et al. 2013). So far 18 species of water bugs are recorded for Montenegro. This is a relatively small number as compared to the neighboring countries (Serbia, Macedonia and Croatia) where more than 50 species are reported in each of them (Aukema, Rieger 1995; Protić 1998; 2011; Kment, Beran 2011; Aukema et al. 2013; Boda et al. 2015).

The objective of the present study was to determine which assemblages of water bugs (Heteroptera: Nepomorpha and Gerromorpha) occur in spring habitats in Montenegro, to evaluate the impact of environmental factors on the spatial pattern of these assemblages and to check the congruence of water bug assemblages based on biotic and environmental classification of karstic spring habitats.

## Study area

Montenegro is the Western Balkan State and covers a total area of 14,026 km<sup>2</sup>. It is a large karstic region and most of the country is covered by the Dinaric Alps. The highest point in the country is Bobotov Kuk (2522 m). The lower areas of Montenegro include the valley of the Zeta River, the Skadar Lake depression and a narrow coastal plain. Biogeographically, Montenegro belongs to the Alpine and Mediterranean

regions which divide the country into two almost equal halves (EEA 2016). The lower areas of Montenegro have a warm Mediterranean climate, with hot and dry summers and cool, rainy winters.

The study was conducted in the area located in the Skadar Lake drainage basin. Lake Skadar is the largest lake in the Balkan Peninsula with a surface area that seasonally fluctuates between 370 to 600 km<sup>2</sup>. There are a number of temporary and permanent karstic springs, some of which are sublacustrine in cryptodepressions (so called 'oko') (Pešić, Glöer, 2013). The majority of springs are vaclusian springs and they are mostly cave springs.

## Materials and methods

Water bugs were sampled with a small Surber sampler (10 × 10 cm = 0.01 m<sup>2</sup>, 350 µm mesh width). The sampling was done in summer 2014. All samples were immediately preserved in 96% ethanol, and subsequently sorted and determined in the laboratory.

In total, 230 adult and larvae specimens of water bugs were collected. The material was identified mostly based on Tamanini (1979) and Macan (1976).

Samples were collected from 32 springs located in the central part of Montenegro (Table 1). At the each site, water temperature and pH were measured with a pH-meter (HI 98127, 0.1 accuracy). Springs were divided into four classes based on their size: 1: <1 m<sup>2</sup>, 2: 1-5 m<sup>2</sup>, 3: 5-20 m<sup>2</sup>, 4: >20 m<sup>2</sup>. Water discharge was determined visually and grouped in classes (Table 2): 1 (<1 l min<sup>-1</sup>), 2: (>1 and <5 l min<sup>-1</sup>), 3: (>5 and <25 l min<sup>-1</sup>), 4: (>25 l min<sup>-1</sup>) according to Von Fumetti et al. (2006). The substrate types were categorized into five classes of frequency (Table 2) based on the percentage cover (Von Fumetti et al. 2006): 0: 0%; 1: 1-25%; 2: 26-50%; 3: 51-75%; 4: 76-100%.

## Data analyses

Statistical analyses were performed using PRIMER 7.0 (Clarke, Gorley 2015) and MVSP v3.21 (Kovach 1998-2012). For cluster analysis based on environmental data, centered and standardized environmental data were classified by the Euclidean distance similarity index. For classification of biotic samples, the Bray-Curtis similarity index on square root transformed data was used. PCA was performed on centered and standardized environmental data of the site groups used in the previous cluster analysis. SIMPER analysis was performed to test differences within faunal composition of groups A, B and C, and I, II, III and IV. Using the SIMPER procedure, dissimilarities

Table 1

## General characteristics of the studied springs

Code	Spring	Longitude (E)	Latitude (N)	Altitude (m)	Spring type	Land use
S1	Skadar Lake area, spring "Karuč"	19°6'20.8"	42°21'29.8"	12	sublacustrine	Lake, village
S2	Skadar Lake area, spring "Sinjac"	19°09'11.8"	42°22'01.6"	9	rheocrene	edge of forest
S3	Podgorica area, spring "Kaludерово oko"	19°8'58.6"	42°22'28.31"	17	sublacustrine	meadows, edge of forest
S4	Podgorica area, Bandići, spring "Cmo oko"	19°9'14.95"	42°29'3.76"	38	limnocrene	village
S5	Podgorica area, Bandići, spring "Vriješko vrelo"	19°10'25.2"	42°29'09.6"	39	limnocrene	village
S6	Podgorica, village Daljam, spring "Kraljičino oko"	19°08'44.2"	42°28'52.3"	44	rheocrene	meadow
S7	Podgorica, spring "Vrela ribnička"	19°17'57.1"	42°26'10.7"	55	rheocrene	urban
S8	Podgorica, Piperi, spring "Studenci"	19°14'35.5"	42°28'59.4"	48	rheocrene	meadow
S9	Podgorica, Piperi, spring "Taban"	19°13'08.5"	42°31'39.3"	89	rheocrene	edge of forest
S10	Danilovgrad, spring „Glava Zete“	18°59'48.7"	42°40'29.5"	78	rheocrene	edge of forest
S11	Danilovgrad, spring „Milojevića vrelo“	19°00'40.3"	42°37'56.1"	50	rheocrene	edge of forest
S12	Danilovgrad, spring „Oraška jama“	19°05'33.1"	42°31'52.5"	56	rheocrene	meadow
S13	Podgorica, Mareza I spring on road to Daljam village	19°10'55.6"	42°28'48.2"	38	limnocrene	rocky ground
S14	Podgorica, Mareza II spring on road to Daljam village	19°10'52.5"	42°28'50.8"	40	rheocrene	meadow
S15	Podgorica, Morača river, spring Zlatica	19°17'18.4"	42°28'07.2"	41	rheocrene	riverside
S16	Podgorica region, Kuči, spring "Mosor"	19°18'32.1"	42°27'46.7"	115	rheocrene (piped)	village
S17	Danilovgrad, village Gornji Martinići, spring "Pištet"	19°11'35.5"	42°33'19.7"	194	rheocrene	village
S18	Danilovgrad, village Gornji Martinići, Glizica, spring	19°10'43.9"	42°33'44.8"	204	rheocrene (piped)	village
S19	Podgorica region, Piperi, spring "Studenac"	19°13'40.7"	42°32'25.2"	443	rheocrene (piped)	village
S20	Podgorica, Piperi, spring "Mrtvak"	19°13'21.2"	42°32'39.7"	406	rheocrene (piped)	village
S21	Podgorica, Piperi, spring "Gospodina voda"	19°13'16.0"	42°32'47.1"	405	rheocrene (piped)	village
S22	Podgorica, Piperi, spring "Bitorod"	19°13'59.9"	42°32'02.3"	404	rheocrene (piped)	village
S23	Podgorica, Piperi, spring "Močila"	19°15'46.6"	42°30'23.9"	106	rheocrene (piped)	edge of forest
S24	Danilovgrad, spring Kupinovo	19°2'47.1"	42°38'35.3"	516	rheocrene (piped)	village
S25	Podgorica region, Kuči, spring "Fundina"	19°21'52.7"	42°26'42.8"	651	rheocrene (piped)	village
S26	Danilovgrad, Podostrog, spring Šobajići	19°2'47.1"	42°38'35.3"	516	rheocrene (piped)	village
S27	Nikšić, village Vidrovan, spring "Vukovo Vrelo"	18°56'31.5"	42°51'26.7"	663	rheocrene	village
S28	Podgorica, spring "Manastir Morača"	19°23'26.1"	42°46'00.4"	308	rheocrene	village
S29	Ponikvica Mt., Martinička Ponikvica, spring I	19°16'01.3"	42°40'28.5"	1419	rheocrene	edge of forest
S30	Lukavica Mt., spring "Babino sicelo"	19°12'54.9"	42°48'15.9"	1607	limnocrene	meadow
S31	Lukavica Mt., spring near Kapetanovo Lake	19°13'40.5"	42°48'46.5"	1706	rheocrene	meadows
S32	Lukavica Mt., spring near Manito Lake	19°14'42.84"	42°48'22.96"	1786	limnocrene	meadow

between and similarities within the above-mentioned groups can be explained with individual species and the composition of Heteroptera assemblages. CCA (ter Braak 1986) was applied to test the influence of environmental variables on the investigated assemblages.

## Results

A total of 25 species were found during this study (Table 3). They represented 9 families. The family Gerridae accounted for 32% (8 taxa) of the total number of taxa, followed by Corixidae (six taxa or 24%), and Notonectidae (four taxa or 16%).

During this study, we found thirteen species new to Montenegro: *Ranatra linearis* (Linnaeus, 1758), *Corixa punctata* (Illiger, 1807), *Hesperocorixa parallela* (Fieber, 1860), *Sigara nigrolineata* (Fieber, 1848),

*S. lateralis* (Leach, 1817), *S. falleni* (Fieber, 1848), *Notonecta maculata* Fabricius, 1794, *N. meridionalis* Poisson, 1926, *Anisops sardeus* Herrich-Schaeffer, 1849, *Aquarius najas* (De Geer, 1773), *A. paludum* (Fabricius, 1794), *Gerris argentatus* Schummel, 1832, and *G. asper* (Fieber, 1860).

From two to ten taxa were found per spring. The maximum  $\alpha$ -diversity (10 species) is found in S2 and S3 – two sublacustrine karst springs in our study. On the other hand, the lowest  $\alpha$ -diversity (2 species) was found in S6. The highest frequency was noted for *Hydrometra stagnorum* (present in 15 springs), followed by *Velia* sp. and *Sigara lateralis*. The most abundant species in the material was *Notonecta glauca*.

According to environmental characteristics, 32 investigated springs can be divided into three groups (Fig. 1A). The results of PCA (Fig. 2A) conducted to determine the environmental patterns most clearly separate the springs from group A. These springs are



Table 2

Physical characteristics (spring size – SU and discharge – DI), temperature (TW), substrate composition and aquatic vegetation of 32 investigated springs (S1-S32)

Spring code	Physical characteristics		TW	Substrate					Aquatic vegetation		
	SU	DI		Anoxic mud	clay	sand	gravel	stones	Moss	Macrophyte	Algae
S1	4	4	17.4	2	1	0	0	1	0	1	1
S2	4	3	18.4	1	1	1	0	1	0	1	1
S3	4	3	18.6	2	1	0	0	1	0	1	1
S4	3	2	15.6	2	1	0	0	1	1	3	1
S5	4	4	14.2	1	1	1	1	1	2	2	2
S6	1	2	13.4	0	0	0	1	2	1	1	1
S7	2	2	12.1	0	1	1	1	2	2	1	1
S8	2	2	14.1	1	1	0	1	2	1	2	1
S9	1	1	16.7	0	1	1	1	1	1	0	1
S10	4	4	11	1	0	0	1	2	3	1	1
S11	3	4	12.3	0	1	1	1	2	2	2	1
S12	4	2	10.1	1	0	1	1	2	1	1	1
S13	3	2	13.1	1	1	1	1	1	1	3	1
S14	2	1	12.4	0	1	1	1	2	1	1	1
S15	2	1	13.1	1	0	2	0	1	2	0	1
S16	2	1	16.1	0	1	0	1	3	1	0	1
S17	1	1	17.2	1	0	0	1	2	1	0	1
S18	1	1	14.3	1	1	0	0	2	1	0	1
S19	1	1	12.3	2	2	0	0	2	1	1	1
S20	2	1	15.2	3	1	0	0	1	0	0	1
S21	2	2	13.5	1	1	1	1	1	1	1	1
S22	1	1	16.4	0	0	1	1	2	1	0	1
S23	1	1	16.1	2	0	0	0	2	1	1	1
S24	1	1	14.1	1	1	0	0	1	1	0	1
S25	1	1	13	1	1	1	1	2	1	0	1
S26	1	1	15.1	2	1	0	0	1	1	0	1
S27	4	4	8.4	1	1	1	1	1	2	2	1
S28	3	4	10.1	1	0	1	1	2	1	1	1
S29	1	1	9.2	2	3	0	0	1	1	0	1
S30	4	1	10.1	3	1	0	0	1	1	2	2
S32	1	1	8.3	3	1	0	0	1	0	1	1
S32	2	1	11.3	3	1	0	0	1	2	2	2

characterized by a high content of anoxic mud, clay and algae, a higher altitude and the lowest concentration of gravel and sand.

The assemblages of water bugs from the investigated springs may be divided into four groups (Fig. 1B). The results of PCA (Fig. 2B) showed that species in group I prefer springs with gravel and sandy substrate, enriched with mosses and characterized by a higher discharge. The communities from group IV generally prefer sites with stony bottom.

Table 4 presents taxa mostly associated with each of the site groups and dissimilarity in the taxonomic composition between each of the groups. *Sigara lateralis* is a characteristic representative of assemblages type I, while *Nepa cinerea* is characteristic of assemblages type II. *Velia* sp. and *Gerris argentatus* are characteristic of springs from group III and IV, respectively. Community groups specified in the biotic site classification (I, II, III, IV) are much better defined,

have higher internal similarity and are more dissimilar to each other than assemblages of A, B and C groups distinguished in the environmental site classification (Table 4).

ANOVA showed significant differences ( $F=5.442$   $p=0.004$ ) for species richness between assemblages of water bugs from site groups I, II, III and IV. Assemblages of type II (mean 7.3, SD 1.0) were characterized by the highest diversity and were followed by assemblages of type III (mean 4.5, SD 1.35), type I (mean 4.0, SD 1.0 and type IV (mean 3.0, SD 1.41).

One-way ANOVA was used to confirm significant differences in species richness ( $F=11.031$   $p=0.000$ ) between the spring types (sublacustrine, limnocrene and rheocrene springs). The LSD analysis revealed that sublacustrine springs significantly differ ( $p=0.000$  and  $p=0.001$ , respectively) from the two other types of springs, while no significant differences were found between rheocrene and limnocrene springs ( $p=0.512$ ).

Table 3

## The occurrence of species in the studied springs

A – abbreviations of the species names. Species new to Montenegro are marked by one asterisk.

Taxa	A	Spring number
Nepomorpha		
Nepidae		
<i>Nepa cinerea</i> Linnaeus, 1758	Nci	1,2,3,4,8,11,12
* <i>Ranatra linearis</i> (Linnaeus, 1758)	Rli	2,3,4,8,27,28
Naucoridae		
<i>Ilyocoris cimicoides</i> (Linnaeus, 1758)	Ici	2,4
Corixidae		
* <i>Corixa punctata</i> (Illiger, 1807)	Cpu	12,18,20,26
<i>Hesperocorixa linnaei</i> (Fieber, 1848)	Hli	1,2,3,27,28
* <i>Hesperocorixa parallela</i> (Fieber, 1860)	Hpa	29,31,32
* <i>Sigara nigrolineata</i> (Fieber, 1848)	Sni	1,3,4,11,17,25
* <i>Sigara lateralis</i> (Leach, 1817)	Sla	1,5,7,10,12,15,21,24,27,28,31,32
<i>Sigara falleni</i> (Fieber, 1848)	Sfa	3,4,12,13,16,19,23,25,30
Notonectidae		
<i>Notonecta glauca</i> Linnaeus, 1758	Ngl	1,5,7,9,11,13,15,18,22,25,31,32
* <i>Notonecta maculata</i> Fabricius, 1794	Nma	3,5,14,16,20,23,25,28,32
* <i>Notonecta meridionalis</i> Poisson, 1926	Nme	1,2,7,14
* <i>Anisops sardeus</i> Herrich-Schaeffer, 1849	Asa	2
Pleidae		
<i>Plea minutissima</i> Leach, 1817	Pmi	1,2,3
Belostomatidae		
<i>Lethocerus patruelis</i> (Stål, 1854)	Lpa	1,3,6,7
Gerromorpha		
Hydrometridae		
<i>Hydrometra stagnorum</i> (Linnaeus, 1758)	Hsa	1,2,4,8,10,13,14,15,17,22,26,27,29,31,32
Veliidae		
<i>Velia</i> sp.	Vaf	9,15,16,18,19,20,21,23,24,25,26,28,30
Gerridae		
* <i>Aquarius najas</i> (De Geer, 1773)	Ana	1, 3, 5, 11, 13, 15, 17, 19, 24, 25
* <i>Aquarius paludum</i> (Fabricius, 1794)	Apa	11, 18, 19, 22, 27, 29, 30, 31
<i>Gerris lacustris</i> (Linnaeus, 1758)	Gla	21, 27, 28, 31, 32
* <i>Gerris argentatus</i> Schummel, 1832	Gar	13, 14, 15, 24, 27
<i>Gerris thoracicus</i> Schummel, 1832	Gth	9, 19
<i>Gerris costae</i> (Herrich-Schaeffer, 1850)	Gco	2, 3
<i>Gerris odontogaster</i> (Zetterstadt, 1828)	God	15, 16, 17, 20, 22, 24
* <i>Gerris asper</i> (Fieber, 1860)	Gas	4, 7, 10

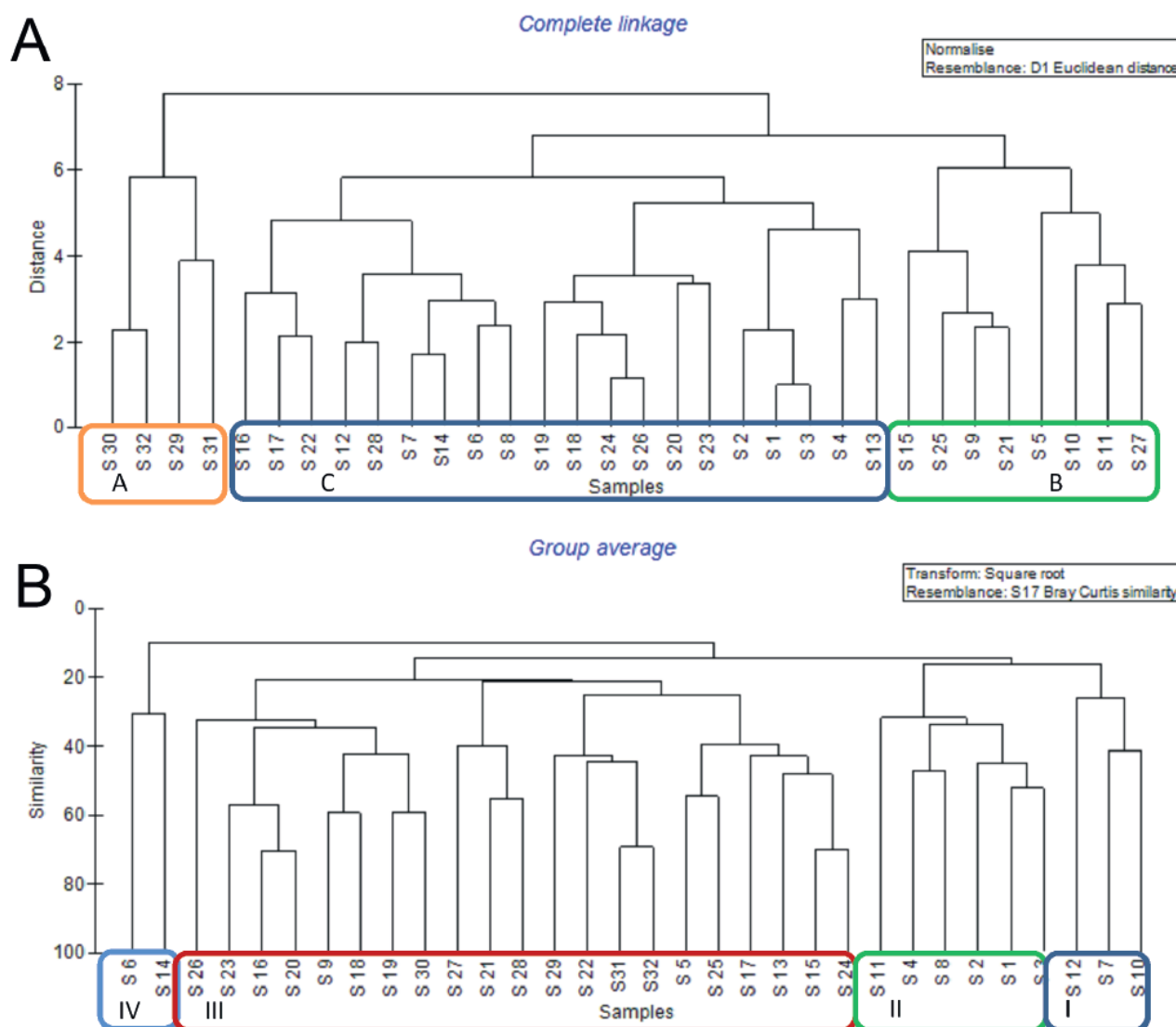
The results of CCA analysis (Fig. 3) summarize the main relationship between aquatic bugs and the environment. Axis 1 and axis 2 explained 25.5% and 18.8% of the total variance, respectively. The environmental factors which significantly influence the communities are spring size (explaining 53.88% and 2.16% of the variation on Axis 1 and 2, respectively), discharge (41.73% and 8.76%, respectively) and water temperature (26.73% and 35.52%, respectively).

Axis 1 was positively correlated with the percentage of mosses and algae and negatively correlated with the spring size, discharge and the percentage of macrophytes. Axis 2 was positively correlated with the altitude, and negatively correlated with the water temperature (Fig. 3).

## Discussion

Aquatic and semiaquatic Heteroptera are an important component of the spring biocenosis. With some exceptions (e.g. Grandova 2014), however, species composition and spatial patterns of water bugs in the springs have not been previously researched. A total of 25 taxa were recorded in 32 springs situated in the central part of Montenegro. This level of diversity is comparable to that reported by Grandova (2014) in springs of the Ukrainian steppe zone (20 species).

The result of our study showed that the fauna of water bugs in the investigated springs is relatively diverse but the widespread polytopic species (*H. stagnorum*, *N. glauca* and *S. lateralis*) dominated.

**Figure 1**

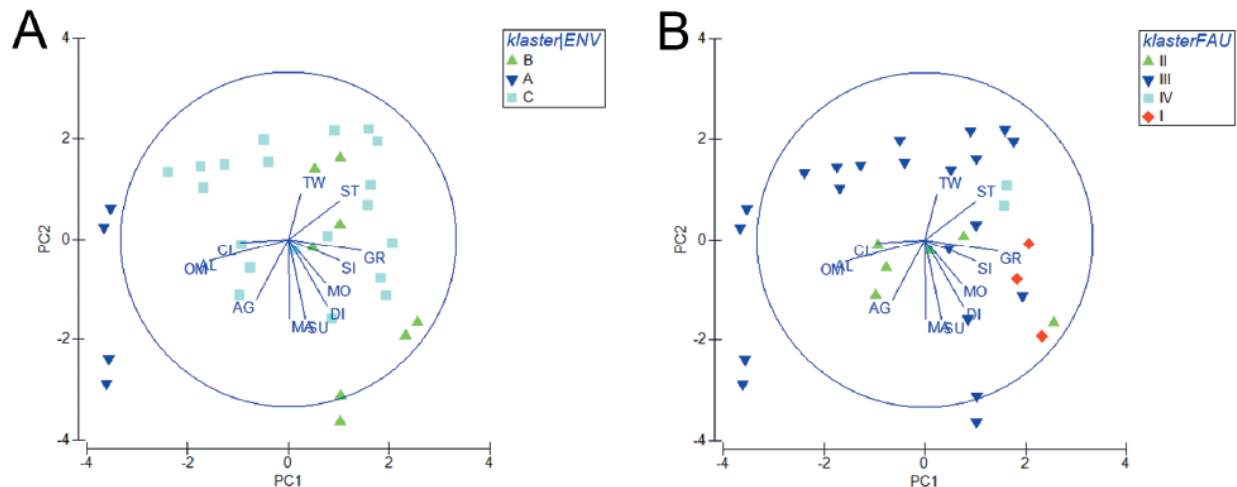
(A) Similarity distance between the sites in groups A, B, and C reflecting environmental characteristics of the investigated springs. (B) Bray-Curtis similarity of water bug assemblages within the investigated springs

However, we could observe some differences in the distribution maxima in different types of spring habitats, even with those common species. *Sigara lateralis* dominated in the assemblages of type I, where 74.52% of the individuals of this species were observed.

Our study in the central part of Montenegro revealed that environmental and faunistic classification of springs based on water bug communities may not correspond with each other. According to environmental characteristics, springs were divided into three groups, while communities of aquatic bug species divide the sites into four groups. The environmental classification indicates anthropogenic impact on the

spring habitats. Group C includes springs which are under a strong anthropogenic influence, especially springs S18-S20 and S22-S24 which are piped (spring water emerging from an artificial pipe) and situated in a village. Springs of group A are located at a higher altitude on the muddy bottom, and their water is not used intensively. Olosutean & Ilie (2010b) proved that the lack of anthropogenic activities favors a heterogeneous community of aquatic bug species.

Our study showed that community groups of water bugs specified in the biotic classification of spring habitats are much better defined than the assemblages distinguished in the environmental



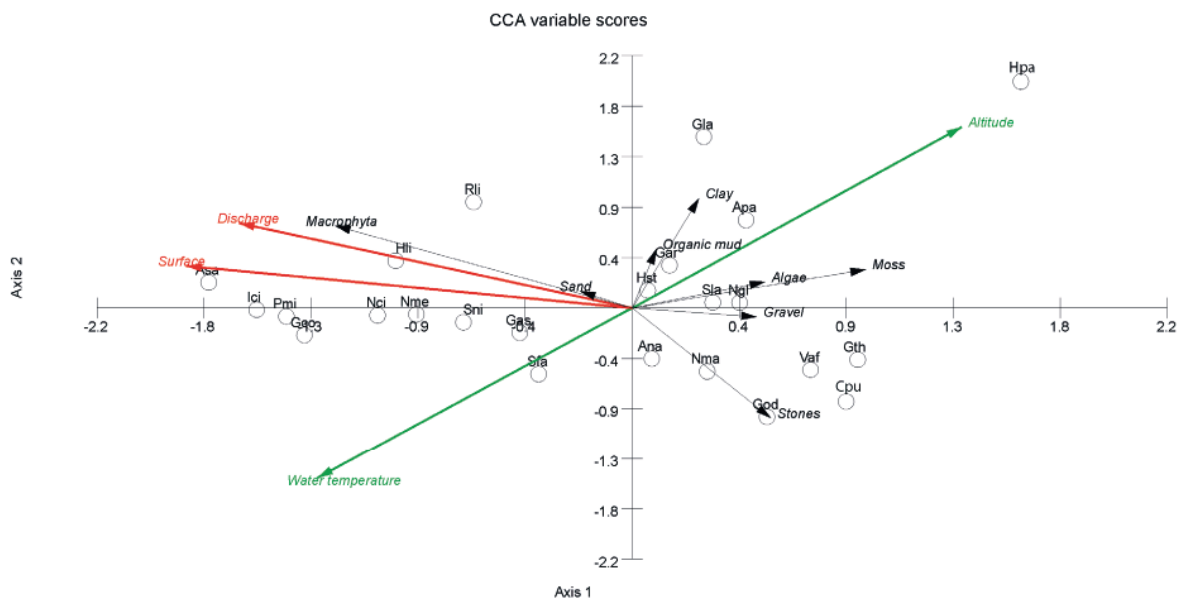
### Figure 2

(A) Results of PCA showing environmental characteristics of 32 investigated springs in relation to environmental classification into groups A, B and C. (B) Results of PCA showing environmental characteristics of 32 investigated springs in relation to faunistic classification into groups I, II, III and IV. Abbreviations: TW – water temperature, DI – discharge, SU – spring size, AL – altitude, ST – stones, GR – gravel, SI – sand, CL – clay, OM – organic mud, AG – algae, MO – moss, MA – macrophytes

site classification. It is worth mentioning that our study focused on the factors directly affecting the water bugs in the aquatic environment. However, the formation of water bug assemblages is also affected by other factors (e.g. the type and structure of landscape, geographical location and proximity of nearby sources

of immigrants) acting in the terrestrial environment and this can cause a discrepancy in the grouping of sites based on the environmental and faunistic data.

Assemblage I dominated by corixid *Sigara lateralis* seems to be characteristic of large karstic lowland springs (S 12, S7, S10) with stony substrate, enriched



### Figure 3

CCA biplot of species and environmental variables based on 32 investigated springs



Table 4

Results of SIMPER analysis for aquatic bug assemblages of site groups A, B and C, and site groups I, II, III and IV

Group A average similarity: 28.98			Groups A and B average dissimilarity = 78.02			Group I average similarity: 29.07			Group I and II average dissimilarity = 88.21		
Group B average similarity: 20.41			Groups A and C average dissimilarity = 85.97			Group II average similarity: 32.48			Group I and III average dissimilarity = 86.72		
Group C average similarity: 15.43			Groups B and C average dissimilarity = 81.84			Group III average similarity: 23.25			Groups I and IV average dissimilarity = 91.10		
						Group IV average similarity: 26.67			Groups II and III average dissimilarity = 88.70		
									Groups II and IV average dissimilarity = 93.80		
Species	Av. Abund.	Av. Sim	Groups III and IV average dissimilarity = 92.33	Contrib. %	Cum %	Species	Av. Abund.	Av. Sim	Sim/SD	Contrib. %	Cum %
Group A						Group I					
<i>Apa</i>	1.75	11.32	0.90	39.07	39.07	<i>Sla</i>	3.67	21.66	1.22	74.52	74.52
<i>Hpa</i>	2.25	6.15	0.80	21.22	60.29	<i>Gas</i>	1.33	7.41	0.58	25.48	100
<i>Hst</i>	1.25	4.56	0.91	15.75	76.04	Group II					
<i>Ngl</i>	1.75	4.17	0.41	14.38	90.42	<i>Nci</i>	3.67	16.55	2.02	50.96	50.96
Group B						<i>Sni</i>	1.33	3.22	0.70	9.93	60.89
<i>Ngl</i>	2.00	6.69	0.60	32.80	32.80	<i>Rli</i>	1.17	2.39	0.75	7.35	68.24
<i>Sla</i>	1.63	5.05	0.63	24.75	57.55	<i>Hst</i>	1.17	2.21	0.75	6.80	75.04
<i>Vaf</i>	1.13	2.72	0.41	13.35	70.90	<i>Hli</i>	1.67	1.72	0.48	5.28	80.32
<i>Ana</i>	0.88	2.26	0.49	11.10	82.00	<i>Ana</i>	1.17	1.65	0.45	5.07	85.40
<i>Hst</i>	0.75	1.46	0.34	7.17	89.17	<i>Pmi</i>	1.00	0.99	0.45	3.06	88.46
<i>Nma</i>	0.63	0.68	0.19	3.33	92.50	<i>Ici</i>	1.50	0.99	0.26	3.04	91.50
Group C						Group III					
<i>Vaf</i>	1.00	2.86	0.38	18.52	18.52	<i>Vaf</i>	1.48	6.79	0.65	29.20	29.20
<i>Hst</i>	0.95	1.93	0.41	12.52	31.03	<i>Hgl</i>	1.43	4.00	0.43	17.21	46.41
<i>Ana</i>	0.80	1.39	0.28	9.03	40.07	<i>Hst</i>	0.95	2.24	0.40	9.65	56.06
<i>Sfa</i>	0.85	1.38	0.32	8.93	48.99	<i>Ana</i>	0.76	1.78	0.30	7.66	63.73
<i>Nma</i>	0.80	1.14	0.27	7.40	56.39	<i>Sla</i>	0.76	1.73	0.35	7.46	71.19
<i>Nci</i>	1.00	0.95	0.26	6.18	62.56	<i>Apa</i>	0.76	1.60	0.31	6.88	78.06
<i>Ngl</i>	0.70	0.93	0.22	6.01	68.58	<i>Nma</i>	0.67	1.44	0.31	6.18	84.25
<i>God</i>	0.70	0.91	0.22	5.87	74.45	<i>God</i>	0.81	1.24	0.26	5.32	89.60
<i>Sla</i>	0.70	0.82	0.20	5.34	79.79	<i>Sfa</i>	0.48	0.89	0.26	3.83	93.43
<i>Cpu</i>	0.50	0.58	0.16	3.79	83.58	Group IV					
<i>Gar</i>	0.35	0.54	0.17	3.48	87.06	<i>Gar</i>	2	26.67	/	100	100
<i>Rli</i>	0.45	0.38	0.22	2.44	89.50						
<i>Hli</i>	0.70	0.36	0.17	2.35	91.84						

with moss. According to Hufnagel et al. (1999), the corixids *Sigara falleni* and *S. lateralis* are characteristic of deeper water bodies. On other hand, communities of type II, characterized by the highest species richness, show preference for limnocene and sublacustrine springs enriched with muddy substrate and with moderately dense vegetation. Such water bodies were also characterized by relatively higher water temperature.

The preference by most water bugs for sites enriched with muddy substrate and/or dense shoreline vegetation may be explained by the greater possibility of finding hiding places (Skern et al. 2010). *Nepa cinerea* favors shaded habitats with moderately dense vegetation and submerged branches (Peták et al. 2014). According to Hufnagel et al. (1999), the latter species with *Hydrometra stagnorum* and *Geris lacustris* forms an

eco-group characteristic of small shallow sites, which is confirmed by CCA results of our study. Skern et al. (2010) indicate that *H. stagnorum* and *G. lacustris* have a preference for low macrophyte cover and higher cover of shoreline plants. The stony substrate also favors *Gerris lacustris* (see: Olosutean, Ilie 2010a).

The results of CCA showed that the first axis is most significantly determined by the spring size, indicating that this factor is the main driver of biotic diversity of water bugs in springs. The importance of dimensions of water bodies in the determination of the spatial pattern of water bug communities was mentioned by several authors (e.g. Macan 1954; Hufnagel et al. 1999; Skern et al. 2010). In our study, spring size reveals a strong gradient along the first axis for the species (*Anisops sardeus*, *Ilyocoris cimicoides*, *Plea minutissima*) that prefer larger water bodies. This is consistent with



the studies on water bug communities in lotic and lentic waters, which showed that some species (e.g. *Ranatra linearis*, *Ilyocoris cimicoides*, *Plea minutissima* group) favor large and deep water bodies (Hufnagel et al. 1999).

The variability of assemblages along the second axis is mainly determined by the altitude and temperature. Species richness of water bugs tends to increase from uplands to mountain elevations (Tully et al. 1991). In our study, we did not find a significant correlation between the altitude and species richness of water bugs in the investigated springs. CCA indicates that *Hesperocorixa parallela* is associated with springs at higher elevation situated on the shore of mountain lakes.

PCA showed that water temperature varies within assemblage types. Temperature is often associated with other factors such as spring size and shading by aquatic and/or riparian vegetation and might have indirect impact on water bug communities. Shallow water has a relatively higher temperature, while shading can reduce water temperature (Moosmann et al. 2005).

Our study showed that the species-richest springs were sublacustrine springs, followed by limnocrenes and rheocrenes. In the study by Grandova (2014), limnocrenes were the richest in species, fewer species were found in rheocrenes, while helocrenes were sparsely populated. The results of our study revealed that sublacustrine springs exhibited significantly higher diversity in terms of species richness. However, we did not observed significant statistical differences between limnocrenes and rheocrenes. This may suggest that the type of springs is not an important factor determining the species richness of water bugs in springs or its influence is masked by other factors, including the human impact.

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