

Charophytes in the estuarine Curonian Lagoon: Have the changes in diversity, abundance and distribution occurred since the late 1940s?

by

Zofija Sinkevičienė^{1,*}, Martynas Bučas²,
Raimonda Ilginė², Diana Vaičiūtė²,
Marija Kataržytė², Jolita Petkuvienė²

DOI: [10.1515/ohs-2017-0019](https://doi.org/10.1515/ohs-2017-0019)

Category: **Original research paper**

Received: **July 20, 2016**

Accepted: **October 21, 2016**

¹Laboratory of Flora and Geobotany, Nature Research Centre, Institute of Botany, Žaliojų Ežerų Str. 49, 08406 Vilnius, Lithuania

²Marine Science and Technology Center, Klaipėda University, H. Manto Str. 84, LT 92294, Klaipėda, Lithuania

Abstract

The results of the recent (2014-2015) inventory and three historical datasets (1949-1959, 1960-1980 and 1997-2007) were analyzed in order to track the long-term changes of charophytes in the largest estuarine lagoon of the Baltic Sea. The present species composition of charophytes in the estuarine part of the Curonian Lagoon consisted of 7 species, including tolerant to salinity (*Chara aspera*, *C. contraria*, *C. globularis* and *Nitellopsis obtusa*) and typical brackish-water species (*Chara baltica*, *C. canescens* and *Tolypella nidifica*). The highest congruence of species was between the 1997-2007 and 2014-2015 datasets, which covered respectively eutrophication and post-eutrophication periods. The 1949-1959 dataset (closest to the reference conditions) differed by the absence of typical brackish-water species. The 1960-1980 dataset (the major period of eutrophication) was relatively poor in study sites and species. During the last 6 decades, only tolerant to salinity freshwater species were constant and abundant; only *Nitella mucronata* can be considered as extinct. Recently, *C. contraria* became dominant and widespread. The changes in the charophyte species composition, abundance and distribution can be explained by different intensity of surveys and/or density of study sites, but also by the increased exposure to brackish waters since 1980s and/or recently reduced effect of eutrophication.

Key words: macrophytes, Baltic Sea, long-term changes, eutrophication

* Corresponding author: zofjasin@gmail.com

Introduction

Coastal lagoons are among the most productive but at the same time the most disturbed aquatic ecosystems due to the impact of combined anthropogenic and natural stressors (Kennish, Paerl 2010). Therefore, these water bodies are listed as priority habitats of European importance in Annex I of the Habitat Directive (European Parliament, Council of the European Union 1992). Enclosed water bodies (bays, coastal lagoons, etc.) have undergone drastic changes during the last decades, mainly due to increased human eutrophication (Rönnberg, Bonsdorff 2004; HELCOM 2009). Improvement of the ecological status of such water bodies is central to the European Water Framework Directive (European Parliament, Council of the European Union 2000). The Directive establishes a framework for ecological status assessment in respect of the defined reference (pristine) conditions. Reference conditions for biological parameters could be defined using modelling approaches, historical datasets and existing information on pristine sites or expert judgement. Generally, historical information on the species diversity is the most reliable method of determining the pristine conditions.

Aquatic flora, especially charophytes (*Characeae*), is considered to be an important biological quality element in fresh and brackish water bodies (Kufel, Kufel 2002; Mathieson, Nienhuis 1991; and references therein). The dense beds of charophytes provide shelter, food and substrate for benthic invertebrates, fishes and birds (Dugdale et al. 2006; Schmieder et al. 2006). Data on the species diversity and threats to charophytes in the Baltic Sea were summarized by Schubert & Blindow (2003). During the last decade, the extensive research on brackish submerged vegetation and the search for indicator species have been carried out in the inner coastal waters of southern Baltic (Schubert et al. 2007; Selig et al. 2007a,b; Steinhardt et al. 2009; Steinhardt, Selig 2008; Steinhardt, Selig 2011) and northern Baltic (Appelgren, Mattila 2005; Rosqvist et al. 2010; Hansen, Snickars 2014; Torn et al. 2014). Nevertheless, the status of charophytes and their response to environmental conditions in the largest lagoons of the Baltic Sea is still not fully understood, especially in the long-term perspective and including the most recent data.

Therefore, the aim of this study was to assess the recent (2014-2015) diversity and the distribution of charophytes in the Curonian Lagoon and to compare them with the previous (1949-1959, 1960-1980 and 1997-2007) results in order to reveal possible long-term changes of charophytes. We also discussed the temporal patterns of charophyte distribution

by relating them to changes in environmental factors, potentially caused by human impact and climate change.

Materials and methods

The study area

The Curonian Lagoon with an area of 1584 km² is the largest estuarine lagoon in the Baltic Sea. This shallow (the average depth is 3.8 m) and almost freshwater body is situated in the south-eastern part

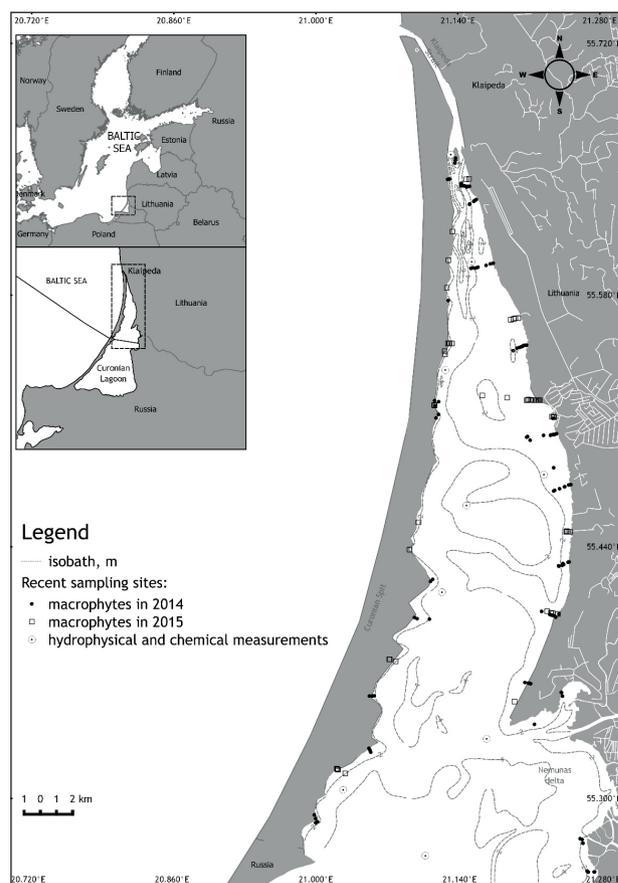


Figure 1

Geographical location of the study area and the distribution of recent (2014-2015) macrophyte sampling transects (samples were collected every ca. 0.25 m along the depth gradient from the coastline or reed stands down to the maximum colonization depth of macrophytes), sites locations (e.g. between the transects or within reed stands) and sampling locations of hydrophysical and chemical parameters in the estuarine part of the Curonian Lagoon

of the Baltic Sea (Zemlys et al. 2013). The lagoon was formed ca. 5000-4000 years ago, when the Curonian Spit separated part of the sea and the Nemunas River Delta (Žaromskis 1996). Therefore, the Curonian Lagoon is recognized as a Natura 2000 site, which borders other territories of European importance, and the Curonian Spit – as a UNESCO World Heritage site.

The northern part of the lagoon (from the Nemunas River Delta) is typically defined as a transitional estuarine system with mixed brackish, lagoon and riverine waters (Fig. 1), whereas the southern part is more lacustrine and characterized by relatively closed water circulation (Ferrarin et al. 2008). The Nemunas River brings 96% of the total freshwater runoff. Saline water from the Baltic flows into the lagoon under prevailing westerly winds in summer and autumn. Episodic inflows of brackish water cause irregular rapid salinity fluctuations (within hours and/or days) in the range of 0.1-7 PSU in the northern part of the lagoon. Brackish waters get into the lagoon through the narrow and ca. 10 km long Klaipėda Strait (port area) and flow mainly through a fairway along the Curonian Spit, while fresh river waters usually flow along the eastern shore of the lagoon (Ferrarin et al. 2008; Zemlys et al. 2013). Consequently, the hydraulic circulation in the lagoon is a factor of primary importance for most of the physical and biogeochemical processes, including changes in water salinity, transparency and concentration of nutrients.

The Curonian Lagoon is one of the most highly eutrophic coastal lagoons in the Baltic Sea (Krevs et al. 2007; Zilius et al. 2014). During the period of 1997-2014, the mean concentrations of total phosphorus (P) ranged from 70 to 180 $\mu\text{g l}^{-1}$ and total nitrogen (N) – from 1100 to 1800 $\mu\text{g l}^{-1}$ (Aplinkos apsaugos agentūra 2014). Nearly continuous decline in the mean concentration of total P began in 1997, which decreased from 150 to 70 $\mu\text{g l}^{-1}$. The concentration of total N did not show clear trends.

Recent mapping of charophytes

The distribution of the study sites during the period of July-September 2014-2015 (Fig. 1) was chosen in line with the previous studies (Minkevičius, Pipinys 1959; Trainauskaitė 1978; Plokštienė 2002; Sinkevičienė 2004) in order to evaluate possible temporal changes in species composition and distribution of charophytes. Two different sampling strategies (transects and points) were used for the assessment of macrophyte species diversity, abundance and depth distribution. The transect sampling approach was conducted at the sites with a relatively wide belt of charophytes, especially on relatively long (>300 m)

and gentle slopes, where samples were collected at a depth of every ca. 0.25 m along imaginary transects extending from the coastline or reed stands down to the maximum colonization depth of macrophytes. The point sampling strategy was used at the sites where previous studies were performed or to check certain locations (e.g. between the transects or within reed stands). In total, 27 transects and 12 sites were studied. Submerged plants (at least 3 samples in each depth zone along the transects) were sampled using a double-headed rake by wading in shallow areas, from a boat and by snorkeling in deep zones. The total coverage (%) of charophytes and abundance of each macrophyte species was assessed using the Braun-Blanquet scale (Kent, Coker 1992). The maximum colonization depth of macrophytes was measured using the marked line and echo-sounder (Humminbird 898c SI Combo).

The macrophyte samples were brought in cooling bags to the laboratory for final species identification. Specimens of charophytes were identified using the keys of Hollerbach & Krasavina (1983), Krause (1997), Blindow & Koistinen (2003); vascular plants – according to Lekavičius (1989) and macroalgae – according to Snoeijs & Johansson (2003).

Historical data

The previous studies of charophytes in the Curonian Lagoon were not systematic and quite scattered, therefore this information was divided into three periods represented by major studies (Table 1): 1949-1959 (Minkevičius, Pipinys 1959), 1960-1980 (Trainauskaitė 1978) and 1997-2007 (Plokštienė 2002; Sinkevičienė 2004). In order to compare the occurrence of charophytes during these periods, the species lists of the first and third period were confirmed with the herbarium material (BILAS, WI).

Data analysis

The recent abundance and depth distribution of charophytes was visualized by kite charts and linear interpolation using packages in R 3.2.0 (R Core Team 2015): “aspace” (Bui et al. 2012), “mapprotools” (Bivand et al. 2015), “geosphere” (Hijmans 2015) and “rgdal” (Bivand, Rundel 2015). The recent and previous distribution of charophytes was analyzed by plotting their records in the geographic information system (ArcGIS 10.3.). Significant differences in the averaged maximum colonization depth of dominant charophytes in different study periods (1949-1959, 1997-2007 and 2014-2015) were determined

Table 1

Characteristics of the recent and historical macrophyte surveys in the Curonian Lagoon used for comparison analysis

Study period	Sampling method	Number of transects and sites	Depth range (m)	Geographic extent	Reference
1949-1959	six-toothed hook, inspection of fishing nets	35 sites (26 – littoral part, 4 – middle part, 5 –relevés); 15 littoral sites in the recent study area	0-4	Entire lagoon (except the south-eastern part)	Minkevičius, Pipinys 1959
1960-1980	no data	no data	no data	Eastern Lithuanian part along the Nemunas River Delta and the south-eastern Russian part	Trainauskaitė 1978
1997-2007	six-toothed hook	22 transects and 13 sites (littoral part)	0-2	Lithuanian part	Plokštienė 2002; Sinkevičienė 2004
2014-2015	double-headed rake, snorkeling	27 transects and 12 sites (littoral part)	0-3	Lithuanian part	this study

by nonparametric multiple contrast effects based on global rankings, computed with the Tukey-type test, using the “nparcomp” package (Konietschke et al. 2015) in R.

The long-term patterns of charophyte distribution were related to changes in environmental factors, such as water salinity, Secchi depth and concentration of phosphates. Data on hydrophysical and chemical parameters were selected with respect to the distribution of charophytes in the study area and vegetation period (May-August). These measurements were grouped according to the four study periods of charophytes: 1954-1957 (Jurevičius 1959), 1960-1972 (Vaitkevičienė, Vaitkevičius 1978), 1997-2007 (national water monitoring by the Department of Marine Research of the Environmental Protection Agency – DMREPA) and 2014-2015 (this study and DMREPA). During the first two periods, the parameters were measured annually – at least once in spring and summer at 16-17 sites, whereas in the two later periods – monthly at 10 sites (Fig. 1). According to the previous studies mentioned above, the sites were classified into the northern and central areas and monthly means of each parameter were calculated for each area. In order to compare the means from the four analyzed periods, the statistics such as the median, 25% and 75% quartiles, the minimum and maximum were derived from them and plotted in R.

Results

Recent diversity and distribution of charophytes

Charophytes were found in 17 out of 27 transects and at 3 out of 12 sites. In total, 7 charophyte species were found during our study, including 5 (*Chara aspera*, *C. contraria*, *C. baltica*, *Nitellopsis obtusa*, *Tolypella nidifica*) distributed mainly in the littoral zone, outside the belt of helophytes (Fig. 2) and 2 species (*Chara canescens* and *C. globularis*) only in the belt

of helophytes. All the species were recorded along the eastern shore of the study area, whereas *C. contraria* was also found in the western part.

C. contraria was the most common species recorded in 59% of the transects. The species colonized down to 2 m depth (Fig. 3). Its monodominant communities were found at different depths, mainly within 0.5-1.5 m and only in the north-eastern part of the lagoon (Fig. 4). All identified specimens were fertile – with gametangia and abundant ripe oospores.

C. aspera was less frequent (recorded in 41% of the transects) and preferred shallow waters between 0.25 and 1.25 m (Fig. 3), even though the maximum colonization depth was 1.7 m. Monospecific communities occurred in the north-eastern part of the study area (Fig. 4). All examined specimens of *C. aspera* had numerous gametangia, but ripe oospores were not observed.

N. obtusa (with relative frequency of 30%) was mainly found in the south-eastern part of the study area to a depth of 2 m (Fig. 3 and 4). It was the only charophyte species recorded in the Nemunas River Delta. *N. obtusa* did not form monospecific communities and was usually the only species co-occurring in the communities of charophytes and angiosperms (Fig. 2). The plants often had unusual appearance; most likely they were damaged by waterfowl.

The rarest and less abundant charophytes were *C. baltica* (19%) and *T. nidifica* (7%), which were recorded in the north-eastern part of the study area (Fig. 4). These species as well as *C. canescens* and *C. globularis* also occurred in the reed stands to a depth of 0.1 m.

The densest vegetation of charophytes (total cover 80-100%) was found in the north-eastern coastal part of the study area (Fig. 2) at a depth of 0.25-1.5 m (Fig. 3). In this part, the maximum colonization depth

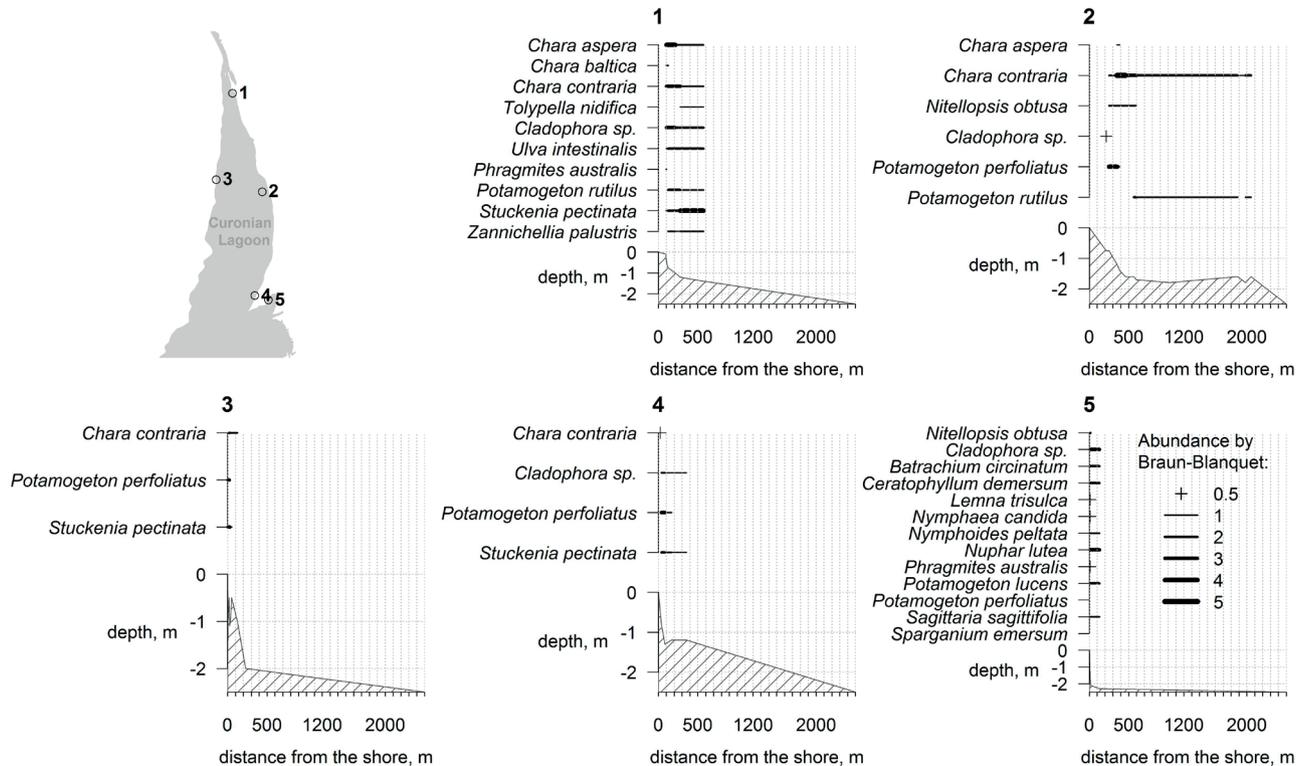


Figure 2

Examples of recent (2014-2015) abundance and depth distribution of macrophytes in the study transects (1-5) with charophytes in the northern part of the Curonian Lagoon

varied between 1.2 and 1.8 m. The number of species and abundance of charophytes decreased toward the south, where the total cover was 5-30% and the maximum colonization depth ranged from 1 to 2 m.

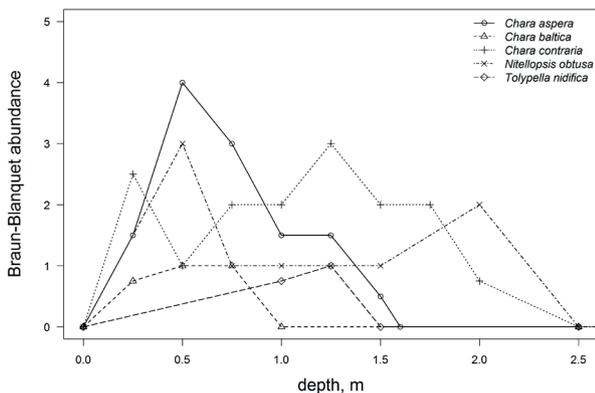


Figure 3

Recent (2014-2015) median ranges of depth distribution and abundance (Braun-Blanquet scale) of 5 charophyte species along the transects (locations are given in Fig. 2) in the Curonian Lagoon; empirical relationships obtained by linear interpolation

Long-term changes of charophytes

Ten charophyte species have been confirmed by the herbarium material since 1949 (Table 2). The first compilation of charophytes in the Curonian Lagoon (Minkevičius, Pipinys 1959) contained 5 species (*Chara aspera*, *C. connivens*, *C. contraria*, *C. globularis* and *Nitellopsis obtusa*). Revision of the herbarium material did not confirm the record of *C. connivens* (Luther 1979; Sinkevičienė 2004), but supplemented the list with *Nitella mucronata*. In the period of 1960-1980, the study sites were scarce, mainly concentrated around the Nemunas River Delta and two new species (*C. virgata* and *C. vulgaris*) were found. The highest diversity of 8 species was recorded during the period of 1997-2007, when 3 new typical brackish-water species (*C. baltica*, *C. canescens* and *Tolypella nidifica*) were found.

Although a different number of species was reported in different periods, *Chara aspera*, *C. contraria* and *Nitellopsis obtusa* were the most constant species during the four study periods. The recent composition of charophyte species was very similar to that in 1997-2007.

The present averaged maximum colonization depth of predominant charophyte species was

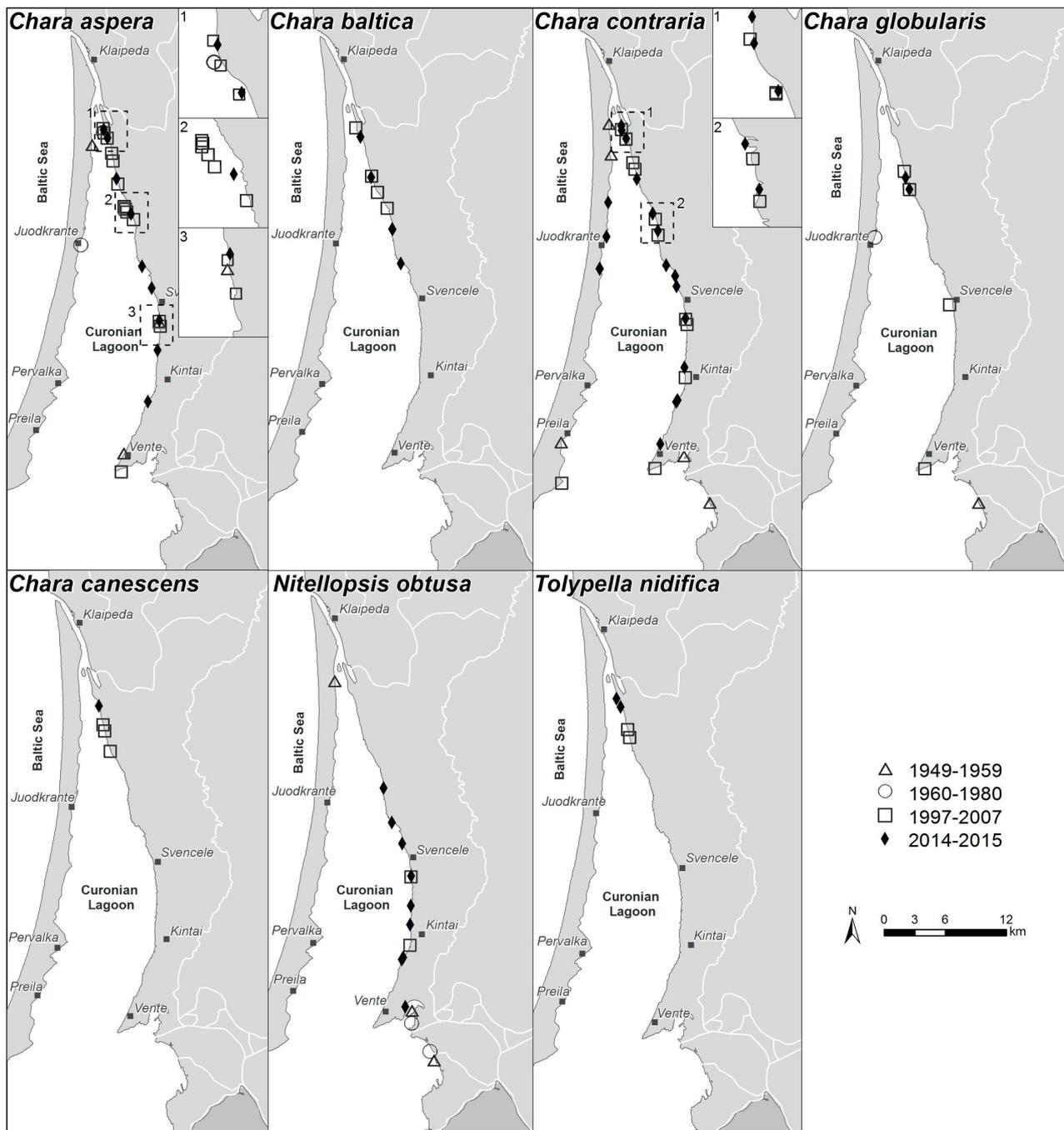


Figure 4

Recent (2014-2015) and previous (1949-1959, 1960-1980 and 1997-2007) distribution of charophyte species recorded in the Curonian Lagoon

statistically significantly (nonparametric Tukey-type contrast $p < 0.001$) higher than in 1949-1959 and 1997-2007 (Table 2), whereas it was significantly higher in 1949-1959 than in 1997-2007.

The data on the distribution of charophytes from the first two periods (1949-1959 and 1960-1980) were quite scarce and could not reflect the whole situation

in the study area, but they provided important information about their occurrence in several localities, especially along the western and south-eastern shores (Fig. 4). *Chara aspera*, *C. contraria*, *C. globularis* and even *Nitellopsis obtusa* were found in the western part of the study area. Moreover, *N. obtusa*, *C. contraria* and *C. globularis* were recorded in the area of the Nemunas

Table 2

The maximum colonization depth (m) and occurrence of charophyte species recorded in different study periods

Species	1949-1959	1960-1980	1997-2007	2014-2015
<i>Chara aspera</i> Willd.	0.8 ^{H,R}	+ ^R	0.6 ^{H,R}	1.7
<i>Chara contraria</i> A. Braun ex Kütz.	0.9 ^{H,R}	+ ^R	0.5 ^{H,R}	2.0
<i>Nitellopsis obtusa</i> (Desv.) J. Groves	0.9 ^{H,R}	+ ^R	0.7 ^{H,R}	2.0
<i>Chara globularis</i> Thuill.	0.9 ^{H,R}	+ ^{H,R}	0.4 ^{H,R}	0.1
<i>Nitella mucronata</i> (A. Braun) Miq.	+ ^H	+ ^H	–	–
<i>Chara vulgaris</i> L.	–	+ ^R	+ ^H	–
<i>Chara virgata</i> Kütz.	–	+ ^{H,R}	–	–
<i>Chara baltica</i> (Hartman) Bruzelius	–	–	0.6 ^{H,R}	0.9
<i>Chara canescens</i> Loisel.	–	–	0.2 ^{H,R}	0.1
<i>Tolypella nidifica</i> (O. F. Müll.) A. Braun	–	–	0.1 ^{H,R}	1.2

H – herbarium data; R – reference data; + – record of a charophyte without maximum colonization depth; – absence of a charophyte; **bolded species** – abundant species

River Delta where they were recently absent. The subsequent studies (1997-2007) revealed the most important sites of charophytes along the eastern shore of the study area and for the first time the localities of typical brackish-water species in the north-eastern part. Recently, *Chara baltica* and *Tolypella nidifica* were observed more widespread in this area.

Discussion

The present study, focused on the charophyte diversity in the largest lagoon of the Baltic Sea, revealed 7 species out of 24 species recorded for the Baltic Sea (Schubert, Blindow 2003). Due to the mixing of freshwater from the Nemunas River and brackish water from the sea, two groups of species could be found at such habitats: freshwater species tolerant to salinity and typical brackish-water species. The similar number of species was revealed earlier in the neighboring Vistula Lagoon (Pliński et al. 1978) and in shallow bays of the northern Baltic Sea (Appelgren, Mattila 2005; Rosqvist et al. 2010). However, the species composition may depend not only on exposure to fresh or brackish water but also other environmental factors such as water transparency, bottom substrate and wave exposure (Kovtun et al. 2011).

Despite the relatively low number of charophyte species recorded in the Curonian Lagoon, we found changes in the species diversity, abundance and distribution during the past 6-7 decades. In the two earlier periods (1949-1959 and 1960-1980), only freshwater species (*Chara aspera*, *C. contraria*, *C. globularis*, *C. vulgaris*, *C. virgata*, *Nitella mucronata* and *Nitellopsis obtusa*) were recorded. The absence of brackish-water species could be explained by the relatively low density of the study sites (Table 1), especially along the north-eastern shore (Minkevičius,

Pipinys 1959). On the other hand, these species could be very rare due to low exposure of brackish waters compared to the later periods (Fig. 5). The increased salinity observed from the 1980s could be explained by enhanced dredging works in the port channel (Gailiūšis et al. 2005) and/or by the change of prevailing wind directions in the eastern Baltic Sea, which increased the occurrence of southwestern waves (Kelpšaitė, Dailidienė 2011) and probably an intrusion of a larger amount of brackish waters compared to 1960-1980. Not surprisingly, the charophytes restricted to brackish environment (*Chara baltica*, *C. canescens* and *Tolypella nidifica*) were recorded for the first time in 1999 (Sinkevičienė 2004) and therefore they were included in the Red Data Book (Sinkevičienė 2007). Only one red-listed species (*N. mucronata*) has not been observed since 1976 and can be treated as extinct (Sinkevičienė 2007).

In the period of 1997-2007, monitoring of macrophytes has begun in the northern part of the Curonian Lagoon (Plokštienė 2002). Several comprehensive surveys of charophytes resulted in the maximum number (8) of species (Sinkevičienė 2004). No significant difference in the species composition was observed in this period compared to 2014-2015. Only one species (*C. vulgaris*) was not found, but it was always a very scarce species that occurred in sheltered bays or reed stands, and therefore could be missed in the recent surveys. Similarly, *Chara canescens* and *C. globularis* were not found in the open coastal areas in 2014-2015, but they were recorded in the reed belt.

In the recent study, a significant dominance and enlargement of areas occupied by *Chara aspera* and *C. contraria* was observed, mainly due to the increase in their maximum colonization depth (Table 2). *C. aspera* was previously treated as a dominant and widespread species, mostly forming monodominant communities in the shallow areas down to 0.5 m depth (Sinkevičienė

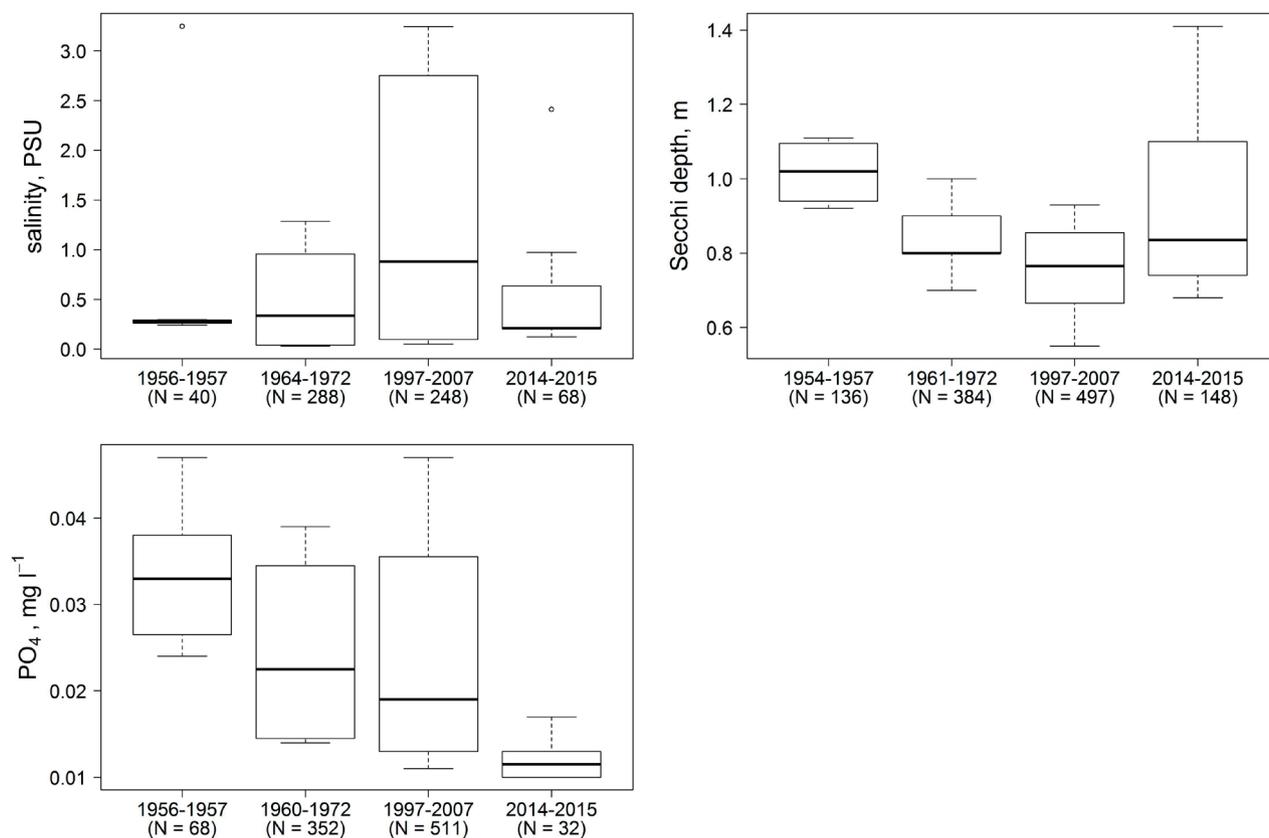


Figure 5

Statistics of mean salinity, Secchi depth and concentration of phosphates (PO_4) in the northern part of the Curonian Lagoon in May-September in different periods: 1954-1957 (Jurevičius 1959), 1960-1972 (Vaitkevičienė, Vaitkevičius 1978), 1997-2007 (DMREPA) and 2014-2015 (this study and DMREPA). The midpoint of a boxplot is the median, the hinges are 25% and 75% quartiles, the whiskers show minimum and maximum values. N – the number of measurements in the period

2004). During the previous periods, *C. contraria* was the only species accompanying *C. aspera*, while recently it formed dense meadows at depths greater than 0.5 m and its maximum colonization depth significantly increased to 2 m. Moreover, the species was recorded at all the sites in 2014-2015 (Fig. 4) where charophytes were found, except 2 study sites in the Nemunas River Delta.

C. contraria is a freshwater species, recorded only occasionally in German, Polish, Lithuanian, Latvian and Estonian coastal waters, in localities with a significant inflow of freshwater (Dekere 2003; Torn et al. 2015). For example, it was frequently recorded in the neighboring Vistula Lagoon (Pliński et al. 1978), in the coastal lakes (e.g. Grosser Binnensee) of the northwestern German Baltic coast, including especially its diaspores in the sediments (Steinhardt, Selig 2011). In the Curonian Lagoon, *C. contraria* was represented by a partially

ecorticate (subgymnophyllous) form, which was also recorded in the Russian part (Romanov, Volodina 2015) and sometimes was treated as a separate species – *C. inconnexa* Allen (Romanov 2015). The recent spread of *C. contraria* in the lagoon could be explained by its relative tolerance to salinity and nutrient enrichment (Dekere 2003). Moreover, *C. contraria* is mentioned as one of the pioneer species colonizing disturbed lakes (Hosper 1994). In the Curonian Lagoon, a decrease in the concentration of nutrients and an increase in water transparency between 1997-2007 and 2014-2015 (Fig. 5) probably enhanced the spread of *C. contraria* in unvegetated areas by means of usually abundant oospores.

C. aspera is the most common species in the Baltic coastal waters, tolerating a wide range of salinity from freshwater to 17-18 PSU (Blindow 2000). It occurs at different depths and is mostly abundant down to 0.5 m

(Nielsen 2003; Torn et al. 2015). Although charophytes were better competitors in clear waters (Blindow 1992), experimental studies evidenced that canopy-forming *Stuckenia pectinata* may outcompete *C. aspera* (Van den Berg et al. 1998). In the Curonian Lagoon, we found the negative relationship between abundance of both species ($r_s = -0.42$, $n = 71$, $p < 0.01$).

Typical brackish-water species – *C. baltica*, *C. canescens* and *Tolypella nidifica* – are frequent species in the Baltic Sea (Blümel 2003; Martin et al. 2003; Urbaniak 2003). In the Curonian Lagoon, they were previously recorded only in reed stands, while *C. baltica* and *T. nidifica* were recently found in the open areas, even below 1 m depth. Their localities were still restricted to the north-eastern part of the lagoon, which is exposed to the inflows of brackish waters. Therefore, the distribution limits of these charophytes corresponded to the modelled annual mean of vertically averaged salinity of 1-2 PSU and exposition time of salinity > 0.5 PSU for 85-150 days (Zemlys et al. 2013). According to Kovtun et al. (2009) and Torn et al. (2004), the distribution of species cover correlated with salinity, shallow depth and sheltered areas, whereas *C. canescens* was mostly found at salinities below 3 PSU in the Estonian coastal waters. In the southern Baltic lagoon (Salzhaff), the diaspore deposition of *C. baltica* and *T. nidifica* was more abundant in sheltered areas (Steinhardt, Selig 2007), supporting the effect of wave exposure.

On the other hand, *Nitellopsis obtusa* distribution in the Curonian Lagoon was associated with riverine waters. *N. obtusa* is a freshwater species, but can be found in brackish waters up to 2.5 PSU (Urbaniak 2003). According to the recent data, the species was mainly distributed in the south-eastern part of the study area, while in the previous periods, it was more widespread in the north-western part (close to the Klaipėda Strait) and in the Nemunas River Delta (Minkevičius, Pipinys 1959; Trainauskaitė 1978). This difference could be explained by lower exposure to brackish waters in the previous periods than thereafter due to the dredging works in the port channel (Gailiūšis et al. 2005) or the change of prevailing wind directions (Kelpšaitė, Dailidienė 2011).

C. connivens is one of the rarest charophyte species in the Baltic Sea (Torn, Martin 2003). Most recently it was rediscovered in the Vistula and Szczecin lagoons (Brzeska et al. 2015), recorded for the first time in the Åland archipelago (Appelgren et al. 2004) and declared as not rare in the Estonian coastal waters (Torn et al. 2004). *C. connivens* was included in the species list of the Curonian Lagoon by Minkevičius & Pipinys (1959). However, after the examination of herbarium specimens from Lithuania, Luther (1979) did

not approve the record (Sinkevičienė 2004). Although we did not find any specimens of *C. connivens* from the Herbaria (BILAS, WI), the presence of this species is very likely in the lagoon. Thus, more thorough surveys of charophyte distribution in the shallow part of the littoral zone are necessary and recommended. According to Selig et al. (2009), there was a relatively high inter-annual variability of the submerged vegetation in the brackish lagoon of the southern Baltic Sea. Therefore, future research should focus not only on abiotic factors, but also on a diaspore bank and the effect of waterfowl community.

In terms of the Water Framework Directive, the first records from 1949-1959 do not cover the pristine conditions (Pardo et al. 2011), because the data on charophytes in the Curonian Lagoon are available only from 1949. However, they are close to the conditions before the major eutrophication occurred (HELCOM 2009). This period is characterized by a relatively high concentration of nutrients and water transparency (Fig. 5). The relatively poor data (the surveys mainly focused on the eastern part of the study area along the Nemunas River Delta) on charophytes from 1960-1980 cover the period of intensive eutrophication since the total load of phosphorus increased 3-4 times and total nitrogen – 5 times during that period (Žaromskis 1996). The lack of information and herbarium material from other parts of the lagoon may be also related to the decline of charophyte vegetation due to eutrophication effects. The period of 1997-2007 can also be classified as a period of intensive eutrophication, because a significant decrease in the concentration of total phosphorus and total nitrogen was observed only from 2010 (Aplinkos Apsaugos Agentūra 2014). The recent study can be attributed to a post-eutrophication period due to a significant decrease in the concentration of nutrients and an increase in water transparency compared to the earlier periods (Fig. 5). This long-term pattern of eutrophication intensity during the analyzed periods most likely affected the changes in the maximum colonization depth of charophytes (and their habitat extent). Despite some temporal differences in the species composition, predominant charophyte species remained almost unchanged throughout the analyzed periods (Table 2). Therefore, the determination of sensitive species is problematic. In other coastal and estuarine parts of the Baltic Sea, the long-term patterns of charophyte composition and distribution, and their relationships with environmental factors are different. For instance, the comparative analysis of charophyte distribution in the Estonian coastal waters (Torn 2004) between similar periods (1970-1978 and 2001) did not show any large-scale changes. According to Kovtun et al. (2009),

however, changes in phytobenthic communities in the Haapsalu Bay over the last 45 years were mainly due to large-scale weather patterns that determined regional salinity and ice conditions, whereas regional nutrient loading had minor effects. On the other hand, eutrophication, reduced pressure of grazers and shores overgrown with reeds were the main causes of long-term changes (between the 1930s and 1940s and 2005/2007) in the distribution of aquatic vascular plants and charophytes in the estuarine area of the Pojoviken-Ekenäs archipelago (Pitkänen et al. 2013).

The datasets for 1997-2007 and 2014-2015 provide comprehensive information about the diversity and distribution of charophytes due to the detailed inventory of species along the shore and transects. This data is valuable for conservation purposes (e.g. continuation of macrophyte monitoring, which was conducted in 1997-2000 and interrupted thereafter, and assessment of the red-listed species status) and the integrated coastal zone management due to the increasing human impact on the lagoon over the last few decades (Jakimavičius, Kovalenkoviėnė 2010). Moreover, changes in the maximum colonization depth of charophytes could be analyzed as one of the water quality assessment elements according to the Water Framework Directive. For instance, the reduced maximum colonization depth of vegetation and the loss of charophyte-dominated plant communities indicate degradation of benthic habitats in the inner coastal waters of the southern Baltic Sea (Selig et al. 2007a). Thus, charophytes are included in the national monitoring program in Germany (Steinhardt et al. 2009) and Estonia (Torn et al. 2014).

Conclusions

The recent charophyte species composition in the estuarine part of the Curonian Lagoon consists of 4 characteristic freshwater species (*Chara aspera*, *C. contraria*, *C. globularis*, *Nitellopsis obtusa*) and typical 3 brackish-water species (*Chara baltica*, *C. canescens*, *Tolypella nidifica*). The species list is similar to the one from 1997-2007, and different compared to the previous periods by the presence of a typical brackish-water species complex.

Despite the increasing water salinity in the lagoon observed from 1949-1959, only freshwater species (*Chara aspera*, *C. contraria*, *Nitellopsis obtusa*) tolerant to salinity were most constant and abundant. Recently, *C. contraria* has become the dominant and more widespread species compared to the previous periods.

The changes in the charophyte species composition, abundance and distribution in the lagoon can

be explained to some extent by different intensity of surveys and density of study sites, but also by temporal patterns of ecological conditions. The presence of typical brackish-water species could be a result of increased exposure to brackish waters due to human activity and/or natural processes. The recent increase of the vegetated area covered by charophytes and their maximum colonization depth could be related to the reduced effect of eutrophication over the last decade. Thus, the data obtained are the basis for further search and testing of water quality indicators in respect of the Water Framework Directive.

Acknowledgements

We wish to thank Dominykas Preibys and Erika Juzėnaitė for the maps. We are grateful to the Department of Marine Research of the Environmental Protection Agency for providing the data of hydrophysical and chemical parameters. Financial support was provided by the Research Council of Lithuania (Contract No. VAT- MIP-040/2014). We are also thankful to reviewers for their comments.

References

- Aplinkos Apsaugos Agentūra. (2014, May). *Ecological status of the transitional and coastal waters in 2014*. Retrieved July 14, 2016, from <http://vanduo.gamta.lt/files/EKOLOGINE%20CHEMIN%C4%96%20B%C5%AAKL%C4%96%202014-05-221432297326713.pdf>. (In Lithuanian).
- Appelgren, K. & Mattila, J. (2005). Variation in vegetation communities in shallow bays of the northern Baltic Sea. *Aquatic Botany* 83: 1-13. DOI: 10.1016/j.aquabot.2005.05.001.
- Appelgren, K., Snickars, M. & Mattila, J. (2004). *Chara connivens* Saltzm. Ex. A. Braun 1835 found in the Åland archipelago – a new species to Finland. *Memoranda Soc. Fauna Flora Fennica* 80: 11-13.
- Bivand, R., Lewin-Koh, N., Pebesma, E., Archer, E., Baddeley, A. et al. (2015). *Maptools: Tools for Reading and Handling Spatial Objects*. R package version 0.8–36. Retrieved July 14, 2016, from <http://CRAN.R-project.org/package=maptools>.
- Bivand, R., Rundel, C. (2015). *rgeos: Interface to Geometry Engine – Open Source (GEOS)*. R package version 0.3-11. Retrieved July 14, 2016, from <http://CRAN.R-project.org/package=rgeos>.
- Blindow, I. (1992). Decline of charophytes during eutrophication: a comparison to angiosperms. *Freshwater Biology* 28: 9-14. DOI: 10.1111/j.1365-2427.1992.tb00557.x.
- Blindow, I. (2000). Distribution of Charophytes along the Swedish Coast in Relation to Salinity and

- Eutrophication. *Internat. Rev. Hydrobiol.* 85: 707-717. DOI: 10.1002/1522-2632(200011)85:5/6<707::AID-IROH707>3.0.CO;2-W.
- Blindow, I. & Koistinen, M. (2003). Determination key for charophytes in the Baltic Sea. In H. Schubert & I. Blindow (Eds.), *Charophytes of the Baltic Sea*. The Baltic Marine Biologists Publication No. 19 (pp. 27-36). Ruggell: A.R.G. Gantner Verlag.
- Blümel, C. (2003). *Chara baltica* Bruzelius 1824. In H. Schubert & I. Blindow (Eds.), *Charophyte of the Baltic Sea*. The Baltic Marine Biologists Publication No. 19 (pp. 53-63). Ruggell: A.R.G. Gantner Verlag.
- Brzeska, P., Woźniczka, A., Pełechaty & M. Blindow, I. (2015). New records of *Chara connivens* P. Salzmänn ex A. Braun 1835 – an extremely rare and protected species in Polish brackish waters. *Acta Societatis Botanicorum Poloniae* 84(1): 143-146. DOI: 10.5586/asbp.2015.010.
- Bui, R., Buliung, R.N. & Rummel, T.K. (2012). *aspace: A collection of functions for estimating centrographic statistics and computational geometries for spatial point patterns*. R package version 3.2. Retrieved July 14, 2016, from <http://CRAN.R-project.org/package=aspace>.
- Dekere, Z. (2003). *Chara contraria* A. Braun ex Kütz. 1845. In H. Schubert & I. Blindow (Eds.), *Charophyte of the Baltic Sea*. The Baltic Marine Biologists Publication No 19 (pp. 89-94). Ruggell: A.R.G. Gantner Verlag.
- Dugdale, T.M., Hicks, B.J., de Winton, M. & Taumoepeau, A. (2006). Fish enclosures versus intensive fishing to restore charophytes in a shallow New Zealand lake. *Aquatic Conservation: Marine and Freshwater Ecosystems* 16: 193-202. DOI: 10.1002/aqc.711.
- Ferrarin, Ch., Razinkovas, A., Gulbinskas, S., Umgieser, G. & Bliūdžiūtė, L. (2008). Hydraulic regime based zonation scheme of the Curonian Lagoon. *Hydrobiologia* 611: 133-146. DOI: 10.1007/s10750-008-9454-5.
- European Parliament, Council of the European Union (1992, May). Council Directive 92/43/EEC of 21 May 1992 on the conservation of natural habitats and of wild fauna and flora. *Official Journal of the European Communities L 206*, 22/07/1992: 0007-0050. Retrieved July 14, 2016, from EUR-Lex: http://eur-lex.europa.eu/legal-content/EN/TXT/?uri=uriserv:OJ.L_.1992.206.01.0001.01.ENG&toc=OJ:L:1992:206:TOC.
- European Parliament, Council of the European Union (2000, December). Directive 2000/60/EC of the European Parliament and of the Council of 23 October 2000 establishing a framework for Community action in the field of water policy. *Official Journal of the European Communities L 327*, 22/12/2000: 0001-0073. Retrieved July 14, 2016, from EUR-Lex: http://www.glassforeurope.com/images/cont/133_26204_file.pdf.
- Gailiūšis, B., Kriaučiūnienė, J. & Kovalenkoviėnė, M. (2005). Studies on permeability of the Klaipėda Strait. In Environmental engineering: the 6th international conference, 26-27 May 2005 (pp. 356-361). Vilnius, Lithuania: Vilnius Gediminas Technical University Press "Technika".
- Hansen, J.P. & Snickars, M. (2014). Applying macrophyte community indicators to assess anthropogenic pressures on shallow soft bottoms. *Hydrobiologia* 738: 171-189. DOI: 10.1007/s10750-014-1928-z.
- HELCOM (2009). Eutrophication in the Baltic Sea – An integrated thematic assessment of the effects of nutrient enrichment and eutrophication in the Baltic Sea region: Executive Summary. *Balt. Sea Environ. Proc.* 115A: 1-148. Retrieved July 14, 2016, from Baltic Sea Environment Proceedings: <http://www.helcom.fi/Lists/Publications/BSEP143.pdf>.
- Hijmans, R.J. (2015). *Geosphere: Spherical Trigonometry*. R package version 1.3-13 [computer software]. Retrieved July 14, 2016, from <http://CRAN.R-project.org/package=geosphere>.
- Hollerbach, M.M. & Krasavina, L.K. (1983). Manual for determination of freshwater algae in USSR. Charophytes. Leningrad: Nauka. (In Russian).
- Hosper, H. (1994). An ecosystem-based approach for the restoration of shallow lakes in the Netherlands. *Lake and Reservoir management* 9: 82.
- Jakimavičius, D. & Kovalenkoviėnė, M. (2010). Long-term water balance of the Curonian Lagoon in the context of anthropogenic factors and climate change. *Baltica* 23(1): 33-46.
- Jurevičius, R. (1959). Hydrochemical characteristic of the Curonian Lagoon. In K. Jankevičius, I. Gasiūnas, A. Gediminas, V. Gudelis, A. Kubickas et al. (Eds.), *The Curonian Lagoon* (Complex research results by Biological Institute of Lithuanian Academy of sciences) (pp. 69-108). Vilnius (In Russian, German and Lithuanian summary).
- Kelpšaitė, L. & Dailidienė, I. (2011). Influence of wind wave climate change to the coastal processes in the eastern part of the Baltic Proper. *Journal of Coastal Research*, SI 64: 220-224.
- Kennish, M.J. & Paerl, H.W. (2010). Coastal lagoons: critical habitats of environmental change. In M.J. Kennish & H.W. Paerl (Eds.), *Coastal lagoons: critical habitats of environmental change* (pp. 1-16). Boca Raton, FL: CRC Press Taylor & Francis Group.
- Kent, M. & Coker, P. (1992). Vegetation description and analysis. London: Belhaven Press.
- Konietschke, F., Placzek, M., Schaarschmidt, F. & Hothorn, L.A. (2015). nparcomp: An R Software Package for Nonparametric Multiple Comparisons and Simultaneous Confidence Intervals. *J. Stat. Softw.*, 64(9), 1-17. DOI: 10.18637/jss.v064.i09.
- Kovtun, A., Torn, K. & Kotta, J. (2009). Long-term changes in a northern Baltic macrophyte community. *Estonian Journal of Ecology* 58(4): 270-285. DOI: 10.3176/eco.2009.4.03.
- Kovtun, A., Torn, K., Martin, G., Kullas, T., Kotta, J. et al. (2011).

- Influence of abiotic environmental conditions on spatial distribution of charophytes in the coastal waters of West Estonian Archipelago, Baltic Sea. *Journal of Coastal Research* 5164: 412-416.
- Krause, W. (1997). *Charales (Charophyceae)*. Freshwater flora of Central Europe (Volume 18). Jena: Gustav Fischer. (In German).
- Kreys, A., Koreiviene, J., Paskauskas, R. & Sulijene R. (2007). Phytoplankton production and community respiration in different zones of the Curonian lagoon during the midsummer vegetation period. *Transitional Waters Bulletin* 1: 17-26. DOI: 10.1285/i1825229Xv1n1p17.
- Kufel, L. & Kufel, I. (2002). Chara beds acting as nutrient sinks in shallow lakes - a review. *Aquatic Botany* 72: 249-260. DOI: 10.1016/S0304-3770(01)00204-2.
- Lekavičius, A. (1989). *Manual for determination of plants*. Vilnius: Mokslas. (In Lithuanian).
- Luther, H. (1979). *Chara connexa* in the Baltic Sea area. *Annales Botanici Fennici* 16: 141-150.
- Martin, G., Torn, K., Blindow, I., Schubert, H., Munsterhjelm, R. et al. (2003). Introduction to charophytes. In H. Schubert & I. Blindow (Eds.), *Charophytes of the Baltic Sea*. The Baltic Marine Biologists Publication No. 19 (pp. 3-14). Ruggell: A.R.G. Gantner Verlag.
- Mathieson, A.C. & Nienhuis, P.H. (Eds.) (1991). Intertidal and Littoral Ecosystems. *Elsevier Science Publication*, Amsterdam.
- Minkevičius, A. & Pipinys, J. (1959). Review of flora and vegetation of the Curonian Lagoon. In K. Jankevičius, I. Gasiūnas, A. Gediminas, V. Gudelis, A. Kublickas et al. (Eds.), *The Curonian Lagoon* (Complex research results by Biological Institute of Lithuanian Academy of Sciences) (pp. 109-137). Vilnius (In Russian, German and Lithuanian summary).
- Nielsen, R. (2003). *Chara aspera* Villd. 1809. In H. Schubert & I. Blindow (Eds.), *Charophyte of the Baltic Sea*. The Baltic Marine Biologists Publication No. 19 (pp. 42-52). Ruggell: A.R.G. Gantner Verlag.
- Pardo, I., Poikane, S. & Bonne, W. (2011). JRC scientific and technical reports. *Revision of the consistency in Reference Criteria application in the phase I of the European Intercalibration exercise*. Luxembourg: Office for Official publications of the European Communities. DOI: 10.2788/27631.
- Pitkänen, H., Peuraniemi, M., Westerbom, M., Kilpi, M. & Numers, M.V. (2013). Long-term changes in distribution and frequency of aquatic vascular plants and charophytes in an estuary in the Baltic Sea. *Annales Botanici Fennici* 50 (SA): 1-54. DOI: 10.5735/085.050.701.
- Pliński, M., Kreńska, B. & Whorowski, T. (1978). Floristic relations and biomass of vascular plants of the Vistula Lagoon. *Stud. Mat. Oceanol.* 21: 161-96. (In Polish).
- Plokštienė, D. (2002). The vegetation structure and peculiarities of its development in the Curonian Lagoon. *Jūra ir aplinka* 2(7): 33-41. (In Lithuanian).
- Romanov, R.E. (2015). *Chara inconnexa* Allen (Streptophyta: Charales) and taxonomic ambiguities associated with subgymnophyllous species close to *C. contraria* A. Braun ex Kütz. s.str. *Cryptogamie, Algologie* 36(4): 371-388. DOI: 10.7872/crya/v36.iss4.2015.371.
- Romanov, R.E. & Volodina, A.A. (2015). State of knowledge of charophytes (Streptophyta: Charophyceae, Charales) in the Kaliningrad oblast (Russia). In Problems of taxonomy and geography of aquatic plants: proceedings of International conference, 21-24 October 2015 (pp. 66-67). Yaroslavl: Filigran. (In Russian).
- Rönnberg, C. & Bonsdorff, E. (2004). Baltic Sea eutrophication: area-specific ecological consequences. *Hydrobiologia* 514: 227-241. DOI: 10.1023/B:HYDR.0000019238.84989.7f.
- Rosqvist, K., Mattilaa, J., Sandström, A., Snickars, M. & Westerbom, M. (2010). Regime shifts in vegetation composition of Baltic Sea coastal lagoons. *Aquatic Botany* 93: 39-46. DOI: 10.1016/j.aquabot.2010.03.002.
- Schmieder, K., Werner, S. & Bauer, H.G. (2006). Submersed macrophytes as a food source for wintering waterbirds at Lake Constance. *Aquatic Botany* 84: 245-250. DOI: 10.1016/j.aquabot.2005.09.006.
- Schubert, H. & Blindow I. (Eds.) (2003). Charophyte of the Baltic Sea. *The Baltic Marine Biologists Publication* No. 19. Ruggell: A.R.G. Gantner Verlag.
- Schubert, H., Schubert, M. & Krause, J.C. (2007). Reconstruction of XIXth century submerged vegetation of coastal lagoons of the German Baltic Sea. *Sea and Environment* 1(14): 16-27.
- Selig, U., Eggert, A., Schories, D., Schubert, M., Blümel, M. et al. (2007a). Ecological classification of macroalgae and angiosperm communities of inner coastal waters in the Southern Baltic Sea. *Ecological Indicator* 7: 665-678. DOI: 10.1016/j.ecolind.2006.07.006.
- Selig, U., Schubert, M., Eggert, A., Steinhardt, T., Sagert, S. et al. (2007b). The influence of sediments on soft bottom vegetation in inner coastal waters of Mecklenburg-Vorpommern (Germany). *Estuarine, Coastal and Shelf Science* 71(1-2): 241-249. DOI: 10.1016/j.ecss.2006.07.015.
- Selig, U., Steinhardt, T. & Schubert H. (2009). Interannual variability of submerged vegetation in a brackish coastal lagoon on the southern Baltic Sea. *Ekológia (Slovak Republic)* 28(4): 412-423. DOI: 10.4149/ekol_2009_04_412.
- Sinkevičienė, Z. (2004). Charophyta of the Curonian Lagoon. *Botanica Lithuanica* 10(1): 33-57.
- Sinkevičienė, Z. (2007). Charophyta. In V. Rašomavičius (Ed.), *Red Data book of Lithuania* (pp. 278-288). Kaunas: Lututė. (In Lithuanian).
- Snoeijs, P. & Johansson, G. (2003). *Swedish Marine and Brackish-Water Algae* (5th ed.). Uppsala University, Uppsala. Retrieved May 16, 2010, from Institutionen för biologisk grundutbildning (IBG), Uppsala University www.ibg.uu.se/1BL170.

- Steinhardt, T. & Selig, U. (2007). Spatial distribution patterns and relationship between recent vegetation and diaspore bank of a brackish coastal lagoon on the southern Baltic Sea. *Estuarine, Coastal and Shelf Science* 74(1): 205-214. DOI: 10.1016/j.ecss.2007.04.004.
- Steinhardt, T. & Selig, U. (2008). Comparison of recent vegetation and diaspore banks along abiotic gradients in brackish coastal lagoons. *Aquatic Botany* 91(1): 20-26. DOI: 10.1016/j.aquabot.2009.01.004.
- Steinhardt, T. & Selig, U. (2011). Influence of salinity and sediment resuspension on macrophyte germination in coastal lakes. *Journal of limnology* 70(1): 11-20. DOI: 10.3274/JL11-70-1-03.
- Steinhardt, T., Karez, R., Weselig, U. & Schubert, H. (2009). The German procedure for the assessment of ecological status in relation to the biological quality element "Macroalgae & Angiosperms" pursuant to the European Water Framework Directive (WFD) for inner coastal waters of the Baltic Sea. *Rostock. Meeresbiolog. Beitr.* 22: 7-42.
- Torn, K. & Martin, G. (2003). *Chara connivens* Salzm. ex A. Braun 1835. In H. Schubert & I. Blindow (Eds.), *Charophyte of the Baltic Sea*. The Baltic Marine Biologists Publication No. 19. (pp.82-88). Rugell: A.R.G. Gantner Verlag.
- Torn, K., Martin, G., Kukk, H. & Trei, T. (2004). Distribution of charophyte species in Estonian coastal waters (NE Baltic Sea). *Scientia Marina* 68: 129-136. DOI: 10.3989/scimar.2004.68s1129.
- Torn, K., Martin, G. & Rostin L. (2014). Testing and development of different metrics and indexes describing submerged aquatic vegetation for the assessment of the ecological status of semi-enclosed coastal water bodies in the NE Baltic Sea. *Estonian Journal of Ecology* 63(4): 262-281. DOI: 10.3176/eco.2014.4.05.
- Torn, K., Kovtun-Kante, A., Herkül, K., Martin, G. & Mäemets, H. (2015). Distribution and predictive occurrence model of charophytes in Estonian waters. *Aquatic Botany* 120: 142-149. DOI: 10.1016/j.aquabot.2014.05.005.
- Trainauskaitė, I. (1978). Aquatic vegetation of the Curonian Lagoon. In K. Jankevičius (Ed.), *Physiological and biochemical terms of planktonic organisms development in the northern part of the Curonian Lagoon* (pp. 61-73). Vilnius: Vaga. (In Russian)
- Urbaniak, J. (2003). *Nitellopsis obtusa* (Desv. in Loisel.) J. Groves 1919. In H. Schubert & I. Blindow (Eds.), *Charophyte of the Baltic Sea*. The Baltic Marine Biologists Publication No. 19. (pp. 99-106). Rugell: A.R.G. Gantner Verlag.
- Vaitkevičienė, O., Vaitkevičius, K. (1978). Hydrochemical conditions. In A. Rainys (Ed.), *The Curonian Lagoon: Hydrological conditions*. (Volume 2). (pp. 81-111). Vilnius: Mokslas. (In Lithuanian).
- Van den Berg, M.S., Coops, H., Simons, J. & de Keizer, A. (1998). Competition between *Chara aspera* and *Potamogeton pectinatus* as a function of temperature and light. *Aquatic Botany* 60(3): 241-250. DOI: 10.1016/S0304-3770(97)00099-5.
- Zemlys, P., Ferrarin, C., Umgieser, G., Gulbinskas, S. & Bellafiore, D. (2013). Investigation of saline water intrusions into the Curonian Lagoon (Lithuania) and two-layer flow in the Klaipeda Strait using finite element hydrodynamic model. *Ocean Science* 9(3): 573-584. DOI: 10.5194/os-9-573-2013.
- Zilius, M., Bartoli, M., Bresciani, M., Katarzyte, M., Ruginis, T. et al. (2014). Feedback mechanisms between cyanobacterial blooms, transient hypoxia, and benthic phosphorus regeneration in shallow coastal environments. *Estuaries and coasts* 37(3): 680-694. DOI: 10.1007/s12237-013-9717-x.
- Žaromskis, R. (1996). *Ocean, Seas, Estuaries*. Vilnius, Debesija. (In Lithuanian).